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EMISSION OF CARBONYL AND POLYAROMATIC HYDROCARBON POLLUTANTS FROM THE COMBUSTION OF LIQUID FUELS: IMPACT OF BIOFUEL BLENDING

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ABSTRACT
The combustion of conventional fuels (Diesel and Jet A-1) with 10-20% vol. oxygenated biofuels (ethanol, 1-butanol, methyl octanoate, rapeseed oil methyl ester, diethyl carbonate, tri(propylene glycol)methyl ether, i.e., CH₃(OC₃H₆)₃OH, and 2,5-dimethylfuran) and a synthetic paraffinic kerosene was studied. The experiments were performed using an atmospheric pressure laboratory premixed flame and a four-cylinder four-stroke Diesel engine operating at 1500 rpm. Soot samples from kerosene blends were collected above a premixed flame for analysis. Polyaromatic hydrocarbons (PAHs) were extracted from the soot samples. After fractioning, they were analyzed by high-pressure liquid chromatography (HPLC) with UV and fluorescence detectors. C₁ to C₈ carbonyl compounds were collected at the Diesel engine exhaust on 2,4-dinitrophenylhydrazine coated cartridges (DNPH) and analyzed by HPLC with UV detection. The data indicated that blending conventional fuels with biofuels has a significant impact on the emission of both carbonyl compounds and PAHs adsorbed on soot. The global concentration of 18 PAHs (1-methyl-naphthalene, 2-methyl-naphthalene, and the 16 US priority EPA PAHs) on soot was considerably lowered using oxygenated fuels, except 2,5-dimethylfuran. Conversely, the total carbonyl emission increased by oxygenated biofuels blending. Among them, ethanol and 1-butanol were found to increase considerably the emissions of carbonyl compounds.

INTRODUCTION
Combustion of conventional fuels such as gasoline, Diesel or jet fuels currently used for transportation significantly contributes to greenhouse emissions and global warming [1]. Recently the interest for synthetic and bio-derived fuels [2-10], which are considered helpful for reducing dependence of both air and ground transportation on petroleum [11-13], and environment friendly, increased. The Fischer-Tropsch process allows the production of kerosene or Diesel types of fuel from synthesis gas. Nevertheless, exhaust emissions released into the atmosphere lead to major environmental and health concerns. Some polluting constituents such as carbon monoxide, nitrogen oxides, unburned hydrocarbons, and particulate matter [14, 15] are strictly regulated by governments. However, other exhaust constituents, generally referred to as unregulated pollutants, have significant impact on atmospheric pollution and human health. Carbonyl compounds (CBCs) and polyaromatic hydrocarbons (PAHs) are among the important unregulated pollutants emitted from internal combustion engines and gas turbines. Carbonyl compounds, i.e., aldehydes and ketones, can be produced from incomplete combustion of fuels [16] and are of significant importance in atmospheric chemistry [17] through their oxidation and interaction with nitrogen oxides which yield other pollutants, e.g., ozone and peroxycetyl nitrates, lowering urban air quality. In addition to their irritative and ecotoxic properties, some carbonyl compounds such as formaldehyde, acetaldehyde and acrolein are identified as probable or known carcinogen and/or mutagen. Polycyclic aromatics originating principally from incomplete combustion of fossil fuels and other organic matter are present on particulates and in the gas phase. PAHs and their derivatives adsorbed on
soot have been largely studied due to their potential carcinogenic and/or mutagenic effects, and reactivity in the troposphere. While there is an increasing interest for developing and using biofuels with the aim of reducing dependency to oil, improving engine efficiency and reducing pollutants formation are needed. Oxygenated and synthetic fuels used pure or as additive are potential alternatives to petroleum-based fossil fuels with expected, but not well-assessed, beneficial effects on emissions.

The purpose of this work is to study the impact on pollutants emissions of blending conventional liquid fuels (Jet A-1 and Diesel) with synthetic jet-fuel and oxygenates that can derive from biomass. The pollutants considered are polyaromatic hydrocarbons, ketones and aldehydes. Carbonyls emissions were measured in the exhaust of a Diesel engine running with Diesel and blends. The identification and quantification of PAHs adsorbed on soot were performed using a premixed flame and Jet A-1, easier to handle with a laboratory burner, as reference fuel. Due to the similar composition of Diesel and Jet A-1 fuels, it is expected that there will be no major impact on the interpretation of the present data.

EXPERIMENTAL

For engine experiments, a conventional Diesel fuel (DF) from Caldeo was used as reference fuel. The other fuels were prepared by blending the DF with 10% vol. of the following oxygenated additives: ethanol (EtNOL), 1-butanol (1-BNOL), methyloctanoate (MOC), diethylcarbonate (DEC), tri(propylene glycol) methyl ether (TPGME), rapeseed oil methyl ester (RME), and a synthetic paraffinic kerosene (SPK) from Sasol having a derived cetane number of 23 (ASTM D7668). Derived cetane numbers of Diesel fuel and fuel blends tested in engine were measured using a Herzog Cetane ID 510 Analyzer (ASTM D7668); they are given in Table 1 together with the average Filter Smoke Number (FSN) which was continuously measured using an AVL-450s smoke meter in the experiments.

Table 1. Fuel characteristics for engine test.

<table>
<thead>
<tr>
<th>Fuel</th>
<th>DCN</th>
<th>FSN</th>
</tr>
</thead>
<tbody>
<tr>
<td>DF</td>
<td>55.34</td>
<td>2.35</td>
</tr>
<tr>
<td>DF/Ethanol*</td>
<td>46.32</td>
<td>2.01</td>
</tr>
<tr>
<td>DF/1-Butanol*</td>
<td>49.71</td>
<td>1.64</td>
</tr>
<tr>
<td>DF/DEC*</td>
<td>48.96</td>
<td>1.15</td>
</tr>
<tr>
<td>DF/MOC*</td>
<td>54.22</td>
<td>2.07</td>
</tr>
<tr>
<td>DF/TPGME*</td>
<td>54.75</td>
<td>1.01</td>
</tr>
<tr>
<td>DF/SPK*</td>
<td>52.74</td>
<td>2.45</td>
</tr>
<tr>
<td>DF/RME*</td>
<td>55.56</td>
<td>1.12</td>
</tr>
</tbody>
</table>

* mixture of Diesel fuel/additive 90:10 v/v

Before refilling the engine with a new fuel, the connection tubing and fuel tank were drained out and the fuel filter was replaced. Prior to each experiment, the engine was allowed to warm up for 30 min to ensure the remaining fuel in the system was consumed, and to stabilize the operating parameters. The experiments were repeated at least once. No after-treatment system was used. Sampling and analysis of CBCs were based on the US EPA TO-11A method [18]. Ten gas samples were collected at 3 m away from the outlet of the combustion chamber by using an automated sampling device (Figure 1). The exhaust gases were collected on 2,4-dinitrophenylhydrazine (DNPH)-coated silica gel cartridges at a constant sampling flow of 0.5 L/min for 5 min. DNPH reacts with carbonyls to form the corresponding stable dinitrophenylhydrazone derivatives that are quantified. At the aforementioned sampling distance from the engine, the gas temperature is low enough (ca. 60°C) to avoid damaging DNPH or the carbonyls derivatives. In order to minimize errors due to the variation of sampling flow rate, ten samples were collected for each fuel, and the average values were used.

The engine used in this study is a direct-injection 4-cylinder 4-stroke Diesel engine with a Continental common-rail injection system. The exhaust gas recirculation rate was set to 32.5%. Experiments were performed at a steady engine operating conditions: BMEP = 0.75 MPa (40% engine load) under a constant speed of 1500 rpm. The main engine characteristics are given in Table 2.

Table 2. Engine characteristics

<table>
<thead>
<tr>
<th>Number of cylinders</th>
<th>4</th>
</tr>
</thead>
<tbody>
<tr>
<td>Bore (mm)</td>
<td>76</td>
</tr>
<tr>
<td>Stroke (mm)</td>
<td>80.5</td>
</tr>
<tr>
<td>Displacement (cm³)</td>
<td>1460.74</td>
</tr>
<tr>
<td>Number of injectors</td>
<td>4</td>
</tr>
<tr>
<td>Compression Ratio</td>
<td>15.21</td>
</tr>
</tbody>
</table>

Figure 1. Schematic representation of the CBCs sampling system used from engine measurements.
After sampling, the cartridges were slowly eluted with 5.0 mL acetonitrile and the extracts were labeled and stored at 4 °C. The samples containing the carbonyl-DNPH derivatives were then analyzed by high-performance liquid chromatography. HPLC grade acetonitrile (ACN), water and tetrahydrofuran (THF) were employed as mobile phase. The following gradient elution was adopted for ACN/water/THF: 0-12 min at 40:55:5 v/v/v and a flow rate of 1.0 mL/min followed by a linear gradient to 55:40:5 v/v/v until 35 min and a flow rate of 1.5 mL/min, and held at 55:40:5 v/v/v until 50 min with a flow rate of 1.5 mL/min. The sample injection volume and column temperature were 20 µL and 40 °C, respectively. The carbonyls detection wavelength was set to 360 nm. The identification and quantification of CBCs were performed using a DNPH-carbonyl standard containing 13 carbonyl compounds: formaldehyde, acetaldehyde, acrolein, acetone, propionaldehyde, crotonaldehyde, methacrolein, 2-butanone, n-butyraldehyde, benzaldehyde, valeraldehyde, p-tolualdehyde, and hexanal.

Table 3. Fuel characteristics for burner experiments.

<table>
<thead>
<tr>
<th>Fuel</th>
<th>O/C in fuel</th>
</tr>
</thead>
<tbody>
<tr>
<td>1: Jet A-1</td>
<td>0.0</td>
</tr>
<tr>
<td>2: SPK</td>
<td>0.0</td>
</tr>
<tr>
<td>3: Jet A-1/1-Butanol</td>
<td>0.040</td>
</tr>
<tr>
<td>4: Jet A-1/DEC</td>
<td>0.092</td>
</tr>
<tr>
<td>5: Jet A-1/MOC</td>
<td>0.040</td>
</tr>
<tr>
<td>6: Jet A-1/2,5-DMF</td>
<td>0.033</td>
</tr>
</tbody>
</table>

For premixed flat-flame burner experiments, a conventional aviation jet fuel (Jet A-1) with a derived cetane number of 44.7 (ASTM D7668) was used as reference fuel. Four fuel blends were prepared by adding 20% by volume of the following oxygenated additives to Jet A-1: 1-butanol, methyl octanoate, diethyl carbonate, 2,5-dimethylfuran, and a synthetic paraffinic kerosene from Sasol. A laboratory flat-flame burner was used for the production and deposition of soot samples from premixed flames of liquid fuels (Jet A-1 and fuel blends) in a reproducible way and under well-characterized and controlled combustion conditions [19]. The soot samples were produced under fuel-rich condition (φ ≈ 2.3) and deposited on the outer surface of a Pyrex tube. The collection tube was introduced perpendicular to the flame axis at the height of 4 cm above the head of the burner and was rotated and moved through the flame by operator. The flame was isolated from environment atmosphere with a cylindrical Pyrex tube through which a nitrogen flow was introduced from the small orifices creating a pressure slightly higher than atmospheric pressure. This avoids the air diffusion and helps to have a homogeneous and stabilized flame. The sampling Pyrex tube was thermostated at 45 °C by an internal water circulation in order to prevent surface temperature gradient and ensure a homogeneous distribution of PAHs along the tube (Figure 2).

![Figure 2. Schematic representation of the apparatus used for soot production and deposition.](image)

The deposited soot was removed mechanically, weighed, and introduced into 2 mL of HPLC-grade n-hexane (Sigma Aldrich with stated purity ≥ 97%). The soot extracts were filtered twice using hydrophobic PTFE filter (pore size 0.2 µm by Alltech) to remove the non-soluble fraction mostly containing carbonaceous material. The resulting soot extracts were analyzed using a Shimadzu HPLC system leading to identification and quantification of PAHs and their derivatives [20]. The HPLC system was equipped with a C18 column (Grace Vydac 201TP, 250 mm × 4.6 mm × 5 µm), an automatic injector, and a photodiode array detector. Acetonitrile and water were used as mobile phase with the elution program: 0-30 min at 50:50 v/v followed by a linear gradient to 100/0 v/v at 90 min, and remained at 100/0 v/v until 110 min. The flow rate and column temperature were 0.5 mL/min and 30 °C, respectively. A certified 18 EPA PAHs standard (2000 µg/mL of each, Supelco) was used for identifying and quantifying naphthalene, 1-methylnaphthalene, 2-methylnaphthalene, acenaphthylene, acenaphthene, fluorene, phenanthrene, anthracene, fluoranthene, pyrene, benzo[a]anthracene, chrysene, benzo[b]fluoranthene, benzo[k]fluoranthene, benzo[a]pyrene, dibenzo[a,h]anthracene, benzo [g,h,i]perylene, and indeno[1,2,3-c,d]pyrene.

RESULTS AND DISCUSSION

Engine results

Thirteen carbonyl compounds were identified and quantified in exhaust emissions of a four-cylinder four-strike Diesel engine running at a stabilized...
operating point with conventional DF and 7 blends (10\% by volume of biofuels or SKP to limit the impact on engine operating conditions such as combustion phasing).

The measured total CBCs emissions are shown in Figure 3. The total concentration of carbonyl compounds obtained for the Diesel fuel was close to 1 mg m$^{-3}$. Based on our results, blending DF with RME and SPK has no major effect on the emission of carbonyl compounds. However, blending DF with ethanol, 1-butanol, diethylcarbonate, methyloctanoate, and TPGME increases significantly the concentrations carbonyl compounds in exhaust gases. Here, the highest CBCs content was measured for the DF/Ethanol blend (3.87 mg m$^{-3}$).

![Figure 3. Total emission of carbonyl compounds.](image)

Carbonyl compounds contribution in exhaust emissions is shown in Figure 4. Under the operating conditions applied in this study, acetaldehyde was the most abundant carbonyl compound measured in the exhausts, for all fuels. Acetaldehyde emissions were in the range 0.32-1.72 mg m$^{-3}$, which corresponds to 33-45\% of the mass of carbonyl compounds in exhaust gases. In addition, the concentration of acetaldehyde in exhaust emissions is higher than formaldehyde for all fuels. Formaldehyde, with a highest contribution of 0.3 mg m$^{-3}$, is the second most abundant carbonyl compound found after acetaldehyde. It represents 6-14\% of the total mass of carbonyl compounds in exhaust gases. The emissions of individual carbonyl compounds are presented in Figure 5.

![Figure 4. Contribution of each carbonyl compound to total emissions of the fuels considered here.](image)

As shown in Figure 5-(a), acetaldehyde is the major component among carbonyls followed by formaldehyde and acrolein. However, one should note that the measured concentrations of carbonyl compounds are very low (less than 1 ppm). A global uncertainty of ± 25\% was estimated. Acetone, propionaldehyde, and crotonaldehyde were also emitted in significant quantities, but their concentrations were below those of the three major compounds cited above. Similar results have been reported by Guarieiro et al. [21] for carbonyl compounds emitted by a Diesel engine operating under similar conditions (1800 or 2000 ppm).

Formaldehyde and acetaldehyde emissions increased for all blended fuels. The highest emissions of formaldehyde (0.223 ppm) and acetaldehyde (0.874 ppm) were measured for the DF/DEC blend and the DF/Ethanol blend, respectively.

Regarding the DF/Ethanol blend, formaldehyde, acetaldehyde and acetone were the most abundant carbonyls. Similar results regarding the increase of formaldehyde and acetaldehyde emissions by adding ethanol to Diesel fuel have been reported by Song et al. [22], and Cheung et al. [23].
Figure 5. Individual carbonyl compounds emissions from the different fuels tested.

The partial combustion of the fuel and the presence of ethanol as the main precursor of acetaldehyde (\(\text{CH}_3\text{CH}_2\text{OH} + \*\text{OH} \rightarrow \text{CH}_3\*\text{CHOH} \rightarrow \text{CH}_3\text{CHO}\)) can explain the high concentrations of formaldehyde and acetaldehyde [24]. As proposed by Tao et al.[25], the degradation of acetaldehyde can result in the formation of \(\text{CH}_3\*\text{CO}\) that can lead to the formation of acetone via:

\[
\text{CH}_3\*\text{CO} + \*\text{CH}_3 \rightarrow \text{CH}_3\text{COCH}_3
\]

Concerning the DF/DEC blend, acetaldehyde was again the most abundant carbonyl emitted, followed by formaldehyde. Such important quantity for acetaldehyde is consistent with experimental results obtained from the oxidation of diethyl carbonate in a jet-stirred reactor [26] where acetaldehyde formation can result from:

\[
(\text{CH}_3\text{CH}_2\text{O})_2\text{C}=\text{O} \rightarrow \text{CO}_2 + \text{C}_2\text{H}_4 + \text{CH}_3\text{CH}_2\text{OH}
\]

\[
\text{CH}_3\text{CH}_2\text{OH} + \*\text{OH} \rightarrow \text{CH}_3\*\text{CHOH} \rightarrow \text{CH}_3\text{CHO} + \text{H}^*
\]

Figure 5-(b) focuses on other carbonyls (methacrolein, 2-butanone, n-butyraldehyde, benzaldehyde, valeraldehyde, p-tolualdehyde and hexanal) which were present in the exhaust emissions but in much lower concentrations. High emissions of n-butyraldehyde and 2-butanone were observed with the DF/1-Butanol blend. The presence of the decomposition products of 1-butanol in the reaction zone could result in the formation of these two carbonyl compounds:

\[
\text{CH}_3(\text{CH}_2)_2\text{OH} + \text{X} \rightarrow \text{CH}_3(\text{CH}_2)_2\*\text{CHOH} + \text{XH}
\]

\[
\text{CH}_3(\text{CH}_2)_2\text{OH} + \text{X} \rightarrow \text{CH}_3\*\text{CH}_2\text{CH}_2\text{CHOH} + \text{XH}
\]

which can in turn yield n-butyraldehyde and 2-butanone. n-Butyraldehyde is formed via:

\[
\text{CH}_3(\text{CH}_2)_2\*\text{CHOH} \rightarrow \text{CH}_3(\text{CH}_2)_2\text{CHO} + \text{H}
\]

The formation of 2-butanone can derive from the low-temperature oxidation of 1-butanol triggered by that of the Diesel fuel (with multiple oxygen addition to fuel-derived radicals, internal hydrogen transfer, and decomposition, followed by recombination of \(\*\text{CH}_3\) and \(\text{CH}_3(\text{CO})\*\text{CH}_2\)):

\[
\text{CH}_3\*\text{CHCH}_2\text{CH}_2\text{OH} (+ 2 \text{O}_2 + \text{CH}_3) \rightarrow 
\text{CH}_3\text{C(O)CH}_2\text{CH}_3 (+ \*\text{OH} + \text{HCOOH})
\]

The seven carbonyls shown in Figure 5-(b) contribute to 10–28% of the total emissions.

**Premixed flame results**

The soot particulates were produced and collected using an atmospheric pressure premixed flame of Jet A-1, SPK, and four biofuel-Jet A-1 blends under fuel-rich conditions (\(\phi \approx 2.3\)). The soluble PAHs fractions were extracted and analyzed by HPLC system using a recently validated analytical methodology [20]. Based on our first observations on premixed flame, the sooting tendency of fuels can be assumed proportional to soot collection rates defined as the total mass of soot particulate deposited in time. One should note that this trend cannot be simply estimated from the O/C ratio in the fuel (Table 3) because it depends strongly on the fuel chemical structure [27-29]. Indeed, blends containing 1-butanol or methyl octanoate produce significantly different amounts of soot whereas they have the same O/C ratio, which is in
line with earlier findings showing alcohols are more efficient to reduce soot than esters [14].

Figure 6. PAHs relative abundance in the analyzed soot extracts. Small PAHs: MW = 128.17-166.22 g/mol; Medium PAHs: MW = 178.23-228.29 g/mol; Large PAHs: MW = 252.31-276.33 g/mol.

The relative sooting tendencies derived here from this assumption were: 1 for Jet A-1, 0.33 for SPK, 0.37 for Jet A-1/1-butanol, 0.50 for Jet A-1/DEC, 1.17 for Jet A-1/Methylloctanoate, and 1.77 for Jet A-1/2,5-dimethylfuran (2,5-DMF). Table 4 shows the measured absolute concentrations of the 16 US EPA priority PAHs present on soot sample. The total PAHs concentration varied from 250 to 4700 ng mg\(^{-1}\).

Table 4. Summary of PAHs measurements.

<table>
<thead>
<tr>
<th>Fuel</th>
<th>Total 18 EPA PAHs (ng mg(^{-1}) of soot)</th>
</tr>
</thead>
<tbody>
<tr>
<td>1 Jet A-1</td>
<td>3551</td>
</tr>
<tr>
<td>2: SPK</td>
<td>3412</td>
</tr>
<tr>
<td>3: Jet A-1/1-Butanol</td>
<td>253</td>
</tr>
<tr>
<td>4: Jet A-1/DEC</td>
<td>1978</td>
</tr>
<tr>
<td>5: Jet A-1/MOC</td>
<td>819</td>
</tr>
<tr>
<td>6: Jet A-1/2,5-DMF</td>
<td>4662</td>
</tr>
</tbody>
</table>

Conventional Jet A-1 and SPK yielded similar PAHs content (3551 and 3412 ng mg\(^{-1}\), respectively). Much higher PAHs content was observed for the Jet A-1/2,5-DMF blend (4662 ng mg\(^{-1}\) of soot). If we apply fuel sooting tendencies on the values presented in Table 4, we can obtain the following trend on total PAHs content for the tested fuel blends: Jet A-1/2,5-DMF > Jet A-1 > SPK > Jet A-1/DEC > Jet A-1/MOC > Jet A-1/1-Butanol. Therefore, excluding the Jet A-1/2,5-DMF blend, the addition of oxygenated biofuels to Jet A-1 reduces considerably the emissions of soot absorbed PAHs.

Table 5. Global toxicity of soot samples*

<table>
<thead>
<tr>
<th>Fuel</th>
<th>Equivalent toxicity (TEQ)*</th>
<th>Variation to Jet A-1 %</th>
</tr>
</thead>
<tbody>
<tr>
<td>Jet A-1/1-Butanol</td>
<td>1.29</td>
<td>-99</td>
</tr>
<tr>
<td>Jet A-1/MOC</td>
<td>83.98</td>
<td>-57</td>
</tr>
<tr>
<td>SPK</td>
<td>115.90</td>
<td>-40</td>
</tr>
<tr>
<td>Jet A-1</td>
<td>193.57</td>
<td>0</td>
</tr>
<tr>
<td>Jet A-1/2,5-DMF</td>
<td>574.14</td>
<td>+197</td>
</tr>
</tbody>
</table>

* calculated global equivalent toxicity according to [30].

The lowest PAHs content was obtained for Jet A-1/1-Butanol blend. Based on our results, the contribution of polyaromatics with different molecular weight (between 125-280 g mol\(^{-1}\)) differs from one soot sample to another. Figure 6 gives the relative abundance of PAHs on soot; the size of the disks is proportional to the quantity of soot produced. Figure 6 shows that light PAHs (with 2 aromatic rings) represents only 3% or less of the total mass of PAHs quantified, while medium size PAHs (with 3-4 aromatic rings) constitute more than 58%. The highest concentration of heavy polyaromatics (≥ 5 aromatic rings) was observed for SPK (lowest sooting tendency) with 42% of the mass of soot collected whereas it reached only 28% of soot collected mass for the Jet A-1/2,5-DMF blend which has the highest sooting tendency.

Global toxicity of soot samples was calculated based on the work of Nisbet and Lagoy [30] that consider the individual toxicity of PAHs. The present work indicates (Table 5) that global toxicity of soot increases when Jet A-1 is blended with 1,2-DMF whereas for other blends, it decreases. Toxicity of soot particulates follows the trend: Jet A-1/2,5-DMF > Jet A-1 > SPK > Jet A-1/MOC > Jet A-1/DEC > Jet A-1/1-Butanol. Such impact of fuel blending, based solely on PAHs content, needs to be confirmed by direct test such as in vitro experiments.

**CONCLUSION**

In this work, we investigated the effect of blending conventional fuels (Diesel and Jet A-1) with oxygenated or synthetic fuels that can be derived from biomass. Carbonyl compounds emission and soot surface-bounded PAHs contents were measured using a four-cylinder Diesel engine and a premixed flame burner, respectively. The results indicated no
correlation between oxygen content of the fuel blends and sooting tendency, PAHs concentrations, or carbonyls emissions. The global concentration of PAHs on soot was considerably lowered using oxygenated fuels, except 2,5-DMF. The lowest PAHs content was found for the 1-butanol-Jet A-1 blend. Conversely, the total carbonyl emission increased when oxygenated biofuels were added to the reference Diesel fuel. Ethanol and 1-butanol were found to increase considerably the emissions of carbonyl compounds. In all samples, acetaldehyde and pyrene were the most abundant compounds among carbonyl and polycyclic aromatic compounds, respectively. This study shows that blending conventional fuels with synthetic fuels does not imply a reduction of pollutants emission nor soot toxicity reduction.

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NOMENCLATURE
φ: Equivalence ratio
BMEP: Brake Mean Effective Pressure
CBs: Carbonyl compounds
DCN: Derived cetane number, ASTM D7668
DF: Diesel fuel
DNPH: 2,4-dinitro phenylhydrazine
HPLC: High pressure liquid chromatography
MW: molecular weight
PAHs: Polyaromatic hydrocarbons
PTFE: Polytetrafluoroethylene
RPM: rotation per minute
SPK: Synthetic paraffinic kerosene
UV: Ultra-violet

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